



Journal of Environmental Sciences 20(2008) 945-951

JOURNAL OF ENVIRONMENTAL SCIENCES ISSN 1001-0742

www.jesc.ac.cn

Dechlorination of disinfection by-product monochloroacetic acid in drinking water by nanoscale palladized iron bimetallic particle

CHEN Chao^{1,2}, WANG Xiangyu^{1,2}, CHANG Ying^{1,2}, LIU Huiling, ^{1,2,*}

1. State Key Laboratory of Urban Water Resources and Environment, Harbin Institute of Technology, Harbin 150090, China. E-mail: chenchao03456@yahoo.com.cn

2. Department of Environmental Science and Engineering, School of Municipal and Environmental Engineering, Harbin Institute of Technology, Harbin 150090, China

Received 11 October 2007; revised 1 November 2007; accepted 18 January 2008

Abstract

Nanoscale palladized iron (Pd/Fe) bimetallic particles were prepared by reductive deposition method. The particles were characterized by X-ray diffraction (XRD), X-ray fluorescence (XRF), scanning electron microscope (SEM), transmission electron microscope (TEM), and Brunauer-Emmett-Teller-nitrogen (BET-N₂) method. Data obtained from those methods indicated that nanoscale Pd/Fe bimetallic particles contained α -Fe⁰. Detected Pd to Fe ratio by weight (Pd/Fe ratio) was close to theoretical value. Spherical granules with diameter of 47 ± 11.5 nm connected with one another to form chains and the chains composed nanoscale Pd/Fe bimetallic particles. Specific surface area of particles was $51 \text{ m}^2/\text{g}$. The factors, such as species of reductants, Pd/Fe ratio, dose of nanoscale Pd/Fe bimetallic particles added into solutions, solution initial pH, and a variety of solvents were studied. Dechlorination effect of monochloroacetic acid by different reductants followed the trend: nanoscale Pd/Fe bimetallic particles of 0.182% Pd/Fe > nanoscale Fe > reductive Fe. When the Pd/Fe ratio was lower than 0.083%, increasing Pd/Fe ratio caused a decrease in DE. Adding more nanoscale Pd/Fe bimetallic particles to solution would enhance DE. The DE of MCAA decreased as initial pH of solution increased.

Key words: nanoscale Pd/Fe bimetallic particles; dechlorination; monochloroacetic acid

Introduction

Haloacetic acids (HAAs) are important disinfection by-products (DBPs) in drinking water. During chlorination of drinking water, large amount of hypochlorous acid (HOCl) is formed in water, meanwhile, some free chlorine reacts with bromide in drinking water to form hypobromous acid (HOBr). Both HOCl and HOBr react with organic matter in drinking water to form DBPs: trihalomethanes (THMs), HAAs, haloacetonitriles (HANs), haloketones (HKs), haloaldehydes (CH), and chloropicrin (CP). It has been reported (Wei and Wang, 2004) that DBPs in drinking water disinfected by liquid chlorine contained 46 wt.% THMs and 42 wt.% HAAs, and DBPs in drinking water disinfected by chloramine contained 24 wt.% THMs and 52 wt.% HAAs. American Water Works Association Research Foundation has reported that unit cancer risk of HAAs (dichloroacetic acid (DCAA): 2.6×10^{-6} , trichloroacetic acid (TCAA): 5.5×10^{-6})) is obviously higher than that of THMs (trichloromethane (TCM): 0.056×10^{-6} , tribromethane (TBM): 0.10×10^{-6} , bromodichloromethane (BDCM): 0.35×10⁻⁶) (Bull and Kopfler, 1991). On the basis of concentration (μg/L) and

* Corresponding author. E-mail: hlliu2002@163.com.

unit cancer risk of DBPs in drinking water, the conclusion that cancer risk (cancer risk of the matter in solution = unit cancer risk concentration) of HAAs is much higher than that of other kinds of DBPs could be obtained. Zhang and Li (2000) have proposed that the concentration of HAAs should be regarded as main parameters to control total cancer risk of DBPs in drinking water. The United States Environmental Protection Agency made regulation about HAAs concentration in drinking water as part of the Disinfectants/Disinfection By-Products Rule (USEPA, 1998), which established the maximum contaminant level for the sum of five HAAs (monochloroacetic acid (MCAA), DCAA, TCAA, bromoacetic acid (MBAA), dibromoacetic acid (DBAA)) of 60 μ g/L. Effective methods on treating HAAs in drinking water have been the focus of the study.

Some researchers (De Wever *et al.*, 2000; Boethling and Alexander, 1979; Yu and Welander, 1995; McRae *et al.*, 2004) conducted studies on biodegradation of HAAs. However, it is hard to use biodegradation method to dechlorinate HAAs in large scope because of long reaction time and strict reaction condition requirements. Therefore, the study on dechlorination of HAAs by zero-valent Fe (Hozalski *et al.*, 2001) is attractive to researchers. Standard electrode potential of Fe²⁺/Fe is -0.44 V, and the electrode potential of Reaction (1) is from +0.5 to +1.5 V, which

indicates that Fe could dechlorinate chlorohydrocarbon (Reaction (1)).

$$RCl + 2e^{-} + H^{+} \longrightarrow RH + Cl^{-}$$
 (1)

Since Gillham and O'Hannesin (1994) proposed that scrap Fe could be used for in situ remediation of groundwater, researchers have conducted large amount of studies on dechlorination of chlorinated organic contaminants with zero-valent Fe. Then, it is found that zero-valent Fe has shortcomings of low reactivity and incomplete dechlorination. To resolve the problems mentioned above, researchers try to use bimetallic particles to dechlorinate contaminants. Ni/Fe (Wei et al., 2004), Cu/Fe (Fennelly and Roberts, 1998), Pt/Fe (Kim and Carraway, 2000; Liu et al., 2000), and Ag/Fe (Xu and Zhang, 2000) are used to treat chlorinated organic contaminants. Muftikian et al. (1995) first use Pd/Fe bimetallic particles to dechlorinate chlorinated organics and the dechlorination efficiency (DE) was greatly enhanced. It is possible to further enhance DE with nanoscale particles because of its high surface area and high activity. Using nanoscale Pd/Fe bimetallic particles, Lien and Zhang (1999, 2001) have dechlorinated chlorinated methanes and chlorinated ethenes, and Xu et al. (2004) have dechlorinated 2,4-dichlorophenol (2,4-DCP).

By studying the mechanisms of dechlorination by Pd/Fe bimetallic particles, researchers (Quan et al., 1998; Cheng et al., 1997; Grittini et al., 1995) propose that: in Pd/Fe bimetallic particles, metal Pd not only serves as catalyst absorbing H₂ but also reduces activation energy of reaction by combining the P electron of chlorine of chlorinated organics or by combining the electron of chlorine with double bonds to form transition complex compound. During dechlorination of chlorinated organics with Pd/Fe bimetallic particles, Fe reacts with H₂O to generate H₂ (Reactions (2) and (3)).

$$Fe + 2H^{+} = Fe^{2+} + H_2 \text{ (in acid solution)}$$
 (2)

$$Fe + 2H_2O = H_2 + Fe^{2+} + 2OH^- \text{ (in alkaline solution)}$$
(3)

The chlorinated organics that form transition complex compounds with Pd and H2 are absorbed by Pd. On the surface of metal Pd, H2 reduces transition complex compounds, removing chlorine from organics to form chlorine ion. Schematic diagram of the mechanism is given in Fig.1.

Hozalski et al. (2001) have studied dechlorination of HAAs by zero-valent Fe with some success. However, MCAA could not be dechlorinated by zero-valent Fe and it existed as product of dechlorination of TCAA. While LD₅₀ for oral-rat of TCAA, DCAA, and MCAA are 400, 2,820, and 580 mg/kg, respectively (Lenga and Votoupal, 1993). This means that dechlorination of TCAA to MCAA could not decrease toxicity of water, and dechlorination of MCAA is necessary. In this study, nanoscale Pd/Fe bimetallic particles were prepared by reductive deposition method and were applied to dechlorinate MCAA in drinking water. Systematic study on dechlorination effect was carried out.

1 Materials and methods

1.1 Preparation of reductants

1.1.1 Preparation of nanoscale Fe

Thirty grams of KBH₄ were added into 179 ml KOH solution (5 g/L) to prepare KBH₄ solution. In an anaerobic box, which was full of nitrogen, nanoscale Fe was produced by adding KBH₄ solution dropwise to 179 ml FeCl₃·6H₂O solution (1 mol/L) (Reaction (4)):

$$4Fe^{3+} + 3BH_4^- + 9H_2O =$$

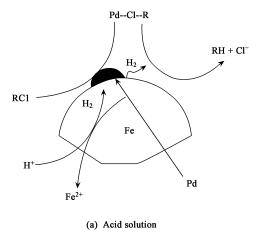
$$4Fe^0 \downarrow + 3H_2BO_3^- + 12H^+ + 6H_2 \uparrow$$
(4)

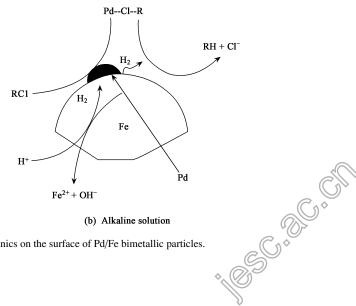
The nanoscale Fe particles were rinsed three times with de-ionized water.

1.1.2 Preparation of nanoscale Pd/Fe bimetallic parti-

Under the protection of nitrogen, nanoscale Fe particles were suspended in ethanol, and then the mixture were charged into a solution of $(Pd(C_2H_3O_2)_2)_3$ in 30 ml of ethanol to generate nanoscale Pd/Fe bimetallic particles according to Reaction (5).

$$Pd^{2+} + Fe^{0} = Pd^{0} + Fe^{2+}$$
 (5)





(b) Alkaline solution

Fig. 1 Schematic of dechlorination of chlorinated organics on the surface of Pd/Fe bimetallic particles.

After being rinsed with dehydrated alcohol and acetone, nanoscale Pd/Fe bimetallic particles were put into vacuum drying oven and heated for 8 h at 100°C. The particles remained in the oven to cool down naturally and then, the particles were ground with one pestle in a mortar. Finally, the particles were preserved in sealed vessels.

1.1.3 Preliminary treatment of reductive Fe

Reductive Fe (98%, 200 mesh) was washed with H₂SO₄ solution (0.1 mol/L) and de-ionized water before reaction. After treatment, the reductive Fe was charged into serum bottle for reaction immediately. And the amount of reductive Fe dissolving into solution by H₂SO₄ washing was analyzed using a 721 spectrophotometer (Shanghai Precision and Scientific Instrument Corporation Limited, China).

1.2 Experiment of dechlorination of MCAA

One hundred milliliters MCAA solution (100 mg/L) and some dose of reductants were added into a 100-ml serum bottle with a rubber stopper packed with Teflon film. The control bottle with MCAA solution only was prepared similarly at the same time. The bottles were oscillated on a platform shaker (170 r/min) at ambient temperature. Samples were taken at stated intervals. And one serum bottle supplied only one sample to avoid the effect of headspace. Also the sample in the control bottle was taken by the same operation. The pH value of solution was adjusted by sulfuric acid and sodium hydroxide. The solution used in this study was prepared by de-ionized water if there was not any special instruction.

1.3 Analytical method

1.3.1 Analytical method for nanoscale Pd/Fe bimetallic particles

The ratio of Pd to Fe by weight of Pd/Fe bimetallic particles was examined by an AXIOSPW4400 X-ray fluorescence (XRF) (PANalytical Company, The Netherlands). Crystal structure was studied by a D/MAX-RB X-ray diffraction (XRD) (Rigaku Company, Japan). Granule size, shape, and surface morphology were observed by a MX2600FE scanning electron microscope (SEM) (Camscan Company, U.K.) and a JEM-1200EX transmission electron microscope (TEM) (JEOL Company, Japan). Specific surface area was detected by an Autosorb-1 Automatic surface area and pore size analyzer (Quantachrome Company, USA). Element content was analyzed by an energy dispersive spectroscope (EDS) (Oxford instruments, U.K.). The amount of Fe in particles was analyzed by a 721 spectrophotometer.

1.3.2 Analytical method for dechlorination of MCAA

Dechlorination effect of MCAA by nanoscale Pd/Fe bimetallic particles was expressed by two parameters, dechlorination efficiency (DE) and generation efficiency (GE). DE was defined as the percentage of metrical chlorine ion in solution to total chlorine element in solution by weight. GE was defined as the percentage of metrical acetic acid (HAc) in solution to HAc generated

from complete dechlorination of MCAA in solution by weight. Concentration of chlorine ion was observed by an ICS 3000 ion chromatograph (IC) (DIONEX Corporation, USA). The quantity of HAc was measured by a HP6890N Gas Chromatograph (GC) (Agilent Company, USA) with flame ionization detection (FID) and a HP-INNDWAX column (30 m \times 0.53 μ m \times 1.0 μ m). The quantity of MCAA was analyzed via HP4890 GC (Agilent Company, USA) with electron capture detection (ECD) and a HP-5 column (30 m \times 0.53 μ m \times 1.0 μ m).

2 Results and discussion

2.1 Characterization of nanoscale Pd/Fe bimetallic particles

2.1.1 Pd/Fe ratio and Fe content of nanoscale Pd/Fe bimetallic particles

The Pd/Fe ratio of nanoscale Pd/Fe bimetallic particles were analyzed by XRF, and the results are shown in Table 1. It can be seen that metrical Pd/Fe ratio was in close agreement with theoretical value.

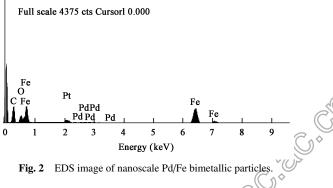
The elements of C, O, Fe, and Pd were detected on the particle surface by EDS (Fig.2). The carbon element might be from ethanol and acetone, which were used to rinse nanoscale Pd/Fe bimetallic particles when particles were prepared. The oxygen element might be from Fe-oxides, ethanol, and acetone. Platinum oxygen was introduced as conductive material coated on specimens when the nanoscale Pd/Fe bimetallic particles were detected by SEM. Spectrophotometer data showed that nanoscale bimetallic particles contained about 90 wt.% Fe.

2.1.2 Crystal structure of nanoscale Pd/Fe bimetallic particles

Crystal structure of nanoscale Pd/Fe bimetallic particles was analyzed by XRD (Fig.3). The d spacings (0.202 nm (110), 0.144 nm (200), and 0.117 nm (211)) from XRD

Table 1 Comparison of theoretical and metrical Pd/Fe ratio

Theoretical Pd/Fe ratio (wt.%)	Metrical Pd/Fe ratio (wt.%)	Standard deviation
0.05	0.051	0.003
0.07	0.067	0.003
0.10	0.083	0.005
0.20	0.182	0.004



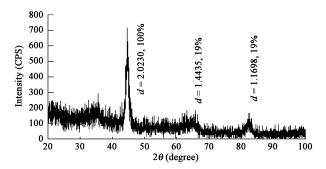


Fig. 3 XRD pattern of nanoscale Pd/Fe bimetallic particles.

pattern were consistent with that for the three strongest reflections of bcc Fe 0 (0.203 nm (110), 0.143 nm (200), and 0.117 nm (211)), suggesting the presence of α -Fe 0 in nanoscale Pd/Fe particles. The Fe $_3$ O $_4$ diffraction peak (0.253 nm (311)) and Fe $_2$ O $_3$ diffraction peak (0.252 nm (311)) were also detected, but the intensity was negligible compared to that of Fe 0 . This indicated that the Fe in nanoscale Pd/Fe bimetallic particles was not oxidized obviously. Pd diffraction peak was not detected because of low Pd/Fe ratio (about 0.05%), and that, it could be detected only when Pd/Fe ratio is higher than 1%.

2.1.3 Surface morphology and granule shape of nanoscale Pd/Fe bimetallic particles

The particles were characterized by SEM and TEM. The results showed that some spherical granules were composed of chains (Fig.4), and some chains connected with each other to form particles with reticular structure (Fig.5). It was difficult to observe the size of particles from SEM and TEM images statistically because of complicate interaction between granules. Figs.4 and 5 show that spherical granules were the basic part of nanoscale Pd/Fe bimetallic particles. This study assumed that the width of

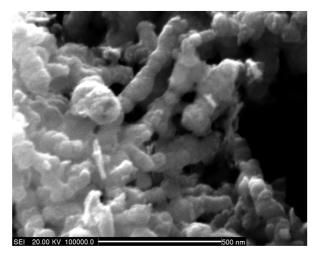


Fig. 4 SEM image of nanoscale Pd/Fe bimetallic particles with magnification of 100,000.

chains in TEM image was consistent with the diameter of spherical granules. And an average granule size of 47 nm with size distribution standard deviation of 11.5 nm was obtained by analyzing 68 spherical granules statistically.

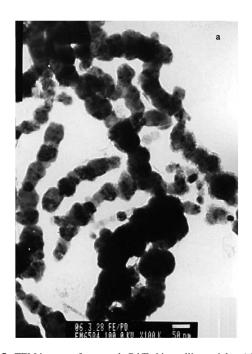
2.1.4 Specific surface area of nanoscale Pd/Fe bimetallic particles

Specific surface area of nanoscale Pd/Fe bimetallic particles was $51 \text{ m}^2/\text{g}$. This specific surface area was much higher than that of Fe and Pd/Fe (0.49 and 0.62 m²/g, respectively) reported in published works (Zhou *et al.*, 2003).

2.2 Dechlorination of MCAA

2.2.1 Dechlorination of MCAA using different reductants

Nanoscale Pd/Fe bimetallic particles, nanoscale Fe, and



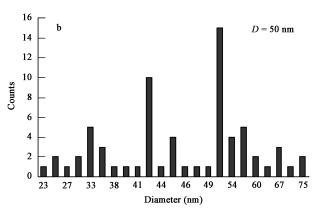


Fig. 5 TEM images of nanoscale Pd/Fe bimetallic particles. (a) magnification of 100,000×; (b) histogram from the TEM image of 68 nanoscale Pd/Fe bimetallic granules.

reductive Fe were used to dechlorinate MCAA, and the results are shown in Fig.6. Decholorination efficiency of different reductants followed the trend: nanoscale Pd/Fe bimetallic particles of 0.182% Pd/Fe ratio > nanoscale Fe > reductive Fe. High specific surface area and high surface activity enabled nanoscale Fe dechlorinate MCAA with a higher DE than reductive Fe with the same reaction time. Further, nanoscale Pd/Fe bimetallic particles possessed a higher DE than nanoscale Fe because nanoscale Pd/Fe bimetallic particles not only had the characteristic of nanoscale Fe of high surface area and high surface activity but also had the characteristic of Pd of facilitating reaction between $\rm H_2$ and chlorinated organics by absorbing these species.

2.2.2 Effect of Pd/Fe ratio of nanoscale Pd/Fe bimetallic particles

Figure 7 shows that when Pd/Fe ratio was less than 0.083%, increasing the Pd/Fe ratio would increase the DE, but when Pd/Fe ratio was more than 0.083%, increasing the Pd/Fe ratio would decrease the DE. Xu and Bhattacharyya (2007) have found that by normalizing surface-area-normalized rate constant ($k_{\rm SA}$) in terms of Pd content, the nanoscale Pd/Fe bimetallic particles with different Pd/Fe ratio studied in their published work possessed the same reaction rate. The result indicated that Pd contents in bimetallic particles played a major role in

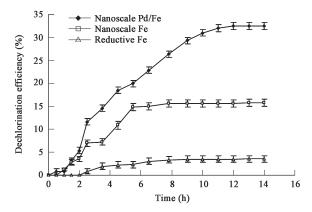


Fig. 6 Dechlorination of MCAA using different reductants.

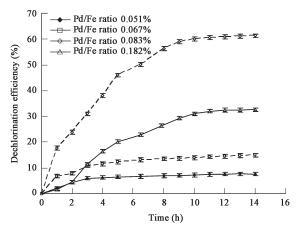


Fig. 7 Effect of Pd/Fe ratio of nanoscale Pd/Fe bimetallic particles on dechlorination of MCAA.

dechlorination of chlorinated organics. The Pd atoms are commonly regarded as surface reactive sites according to Fig.1, and Pd atoms would accelerate H₂ generation rate. In this study, when the Pd/Fe ratio was lower than 0.083%, increasing the Pd/Fe ratio would increase the generation rate of H₂, resulting in increase of instantaneous output of H₂ and DE of MCAA. When Pd/Fe ratio was higher than 0.083%, large amount of H₂ was generated instantly, which was far more than that required for dechlorination of MCAA and H₂ accumulated on nanoscale Pd/Fe bimetallic particle surface. The accumulated H₂ would block mass transfer of MCAA from solution to Fe surface and decrease DE of MCAA.

2.2.3 Effect of dose of nanoscale Pd/Fe bimetallic particles

The results of dechlorination of MCAA with different nanoscale Pd/Fe bimetallic particle dose are shown in Fig.8. After 5 h of reaction, DE of MCAA with different nanoscale Pd/Fe bimetallic particle dose followed the trend (metal to solution ratio) 12 g/L > 10 g/L > 7 g/L > 3 g/L. The reason could be that increasing the amount of nanoscale Pd/Fe bimetallic particles in solution would increase more surface area for reaction between MCAA and Fe, and enhance DE. However, during the first 4 h of reaction, DE with 12 g/L nanoscale Pd/Fe bimetallic particles was lower than that with 10 g/L nanoscale Pd/Fe bimetallic particles. It was possibly because that large amount of H_2 generated from large amount of Fe reduced the mass transfer of MCAA from solution to Fe surface and decreased DE.

2.2.4 Effect of initial pH of solution

The dechlorination efficiency of MCAA with different initial pH is shown in Fig.9. When pH values were 2.98, 7.00, and 9.00, the DE of MCAA were 61.32%, 53.90%, and 43.54%, respectively. As can be seen from Fig.1, dechlorination of MCAA consumed H^+ , and decrease of H^+ concentration would cause DE of MCAA to decrease. Also in alkaline solution, Reaction (6) occurred easily, and the consumption of electron by O_2 would decrease DE of MCAA. However, DE of MCAA with pH 9 was 71% of

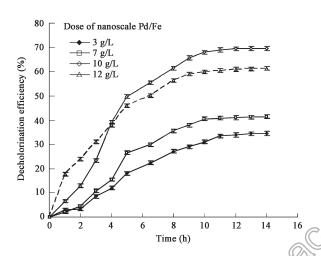


Fig. 8 Effect of dose of nanoscale Pd/Fe bimetallic particles on decidorination of MCAA.

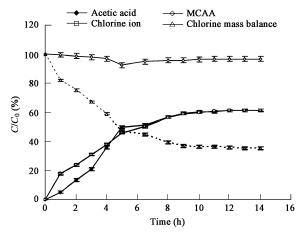


Fig. 9 Effect of solution initial pH on dechlorination of MCAA. Condition: metal to solution ratio, 1 g/100 ml; Pd/Fe ratio, 0.083%.

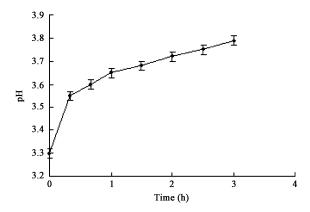


Fig. 10 Concentration change of reactant during dechlorination of MCAA.

that with pH 2.98, indicating the possibility of application of this technology in a larger pH scope. This result could be proved by results reported by other researchers. Wu *et al.* (2004) found that Fe/Cu bimetallic systems dechlorinated CCl₄ well when pH value of solution ranged from 2 to 9, and the lowest DE of CCl₄ was about 75% of the highest DE of CCl₄. Liu and Lowry (2006) studied the effect of pH value on dechlorination of trichloroethene (TCE) with nanoscale Fe particle and found that decreasing pH from 8.9 to 6.5 increased the H₂ evolution rate constant 27 folds and increased the TCE dechlorination rate constant only 2 folds. These results showed that pH did not have great impact on dechlorination of chlorinated organics, and the technology could be applied in a larger pH scope.

$$O_2 + 2H_2O + 4e^- \longrightarrow 4OH^- \tag{6}$$

2.2.5 Concentration of reactant in solution during dechlorination of MCAA

Concentration change of reactant during dechlorination of MCAA was studied. The results from Fig.10 suggested the discrepancy between GE and DE during the first 4 h of dechlorination. The reason could be that nanoscale Pd/Fe bimetallic particles adsorbed some HAc and prolonged the release of HAc to solution. The chlorine mass balance in this study was 96.8%.

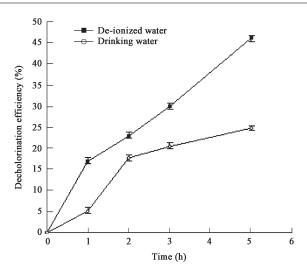


Fig. 11 Dechlorination of MCAA in drinking water and de-ionized water.

2.2.6 Dechlorination of MCAA in drinking water and de-ionized water by Pd/Fe bimetallic particles

Figure 11 shows that DE of MCAA in drinking water was lower than that of MCAA in de-ionized water. It is likely that impurity in drinking water affected the DE of nanoscale Pd/Fe bimetallic particles. Lo et al. (2006) have found that hardness ions and carbonate in groundwater decreased Cr(VI) removal capacity of zero-valent iron by the formation of passivated precipitates on the Fe⁰ surface. It is possible that hardness ions and carbonate were present in drinking water and decreased the DE of MCAA by the same passivated precipitates formed on nanoscale Pd/Fe bimetallic particle surface. Meanwhile, anions such as Cl⁻ and SO₄²⁻ present in water have been found to inhibit dechlorination of TCE with nanoscale zero-valent Fe particles (Liu et al., 2007). Some further research is needed to study the effect of impurity in drinking water on nanoscale Pd/Fe bimetallic particle reactivity with chlorinated organics in detail.

3 Conclusions

The nanoscale Pd/Fe bimetallic particles prepared in this study were characterized. The particles had high specific surface area of 51 m²/g; the diameter of granules was 47±11.5 nm; the particles contained 90 wt.% Fe. Meanwhile, the study on dechlorination of MCAA with nanoscale Pd/Fe bimetallic particles was conducted. The results showed that dechlorination efficiency of MCAA with nanoscale Pd/Fe bimetallic particles was higher than those with nanoscale Fe and reductive Fe. The optimal condition for dechlorination of MCAA was when Pd/Fe ratio of nanoscale Pd/Fe bimetallic particles was 0.083% and particle dose was 12 g/L. During reaction DE of MCAA decreased as solution initial pH increased and DE of MCAA in drinking water was lower than that in deionized water.

Acknowledgements

This work was supported by the Program for the

Changjiang Scholars and Innovation Research Team in University (PCSIRT), National Nature Sciences Foundation of China.

References

- Boethling R S, Alexander M, 1979. Effect of concentration of organic chemicals on their biodegradation by natural microbial communities. *Appl Environ Microbiol*, 37(6): 1211–1216.
- Bull R J, Kopfler F C, 1991. Health Effects of Disinfectants and Disinfection By-Products. Report. AWWARF, Denver, Colorado, USA.
- Cheng F, Fernando Q, Korte N, 1997. Electrochemical dechlorination of 4-chlorophenol to phenol. *Environ Sci Technol*, 31(4): 1074–1078.
- De Wever H, Cole J R, Fettig M R, Hogan D A, Tiedje J M, 2000. Reductive dehalogenation of trichloroacetic acid by *Trichlorobacter thiogenes* gen. nov., sp. nov. *Appl Environ Microbiol*, 66(6): 2297–2301.
- USEPA (United States Environmental Protection Agency), 1998. National Primary Drinking Water Regulations: Disinfectants and Disinfection Byproducts. Final Rule, Federal Register, 241: 69390–69476.
- Fennelly J P, Roberts A L, 1998. Reaction of 1,1,1-trichloroethane with zero-valent metals and bimetallic reductants. *Environ Sci Technol*, 32(13): 1980–1988.
- Gillham R W, O'Hannesin S F, 1994. Enhanced degradation of halogenated aliphatics by zero-valent iron. *Ground Water*, 32(6): 958–967.
- Grittini C, Malcomson M, Fernando Q, Korte N, 1995. Rapid dechlorination of polychlorinated biphenyls on the surface of a Pd/Fe bimetallic system. *Environ Sci Technol*, 29(11): 2898–2900.
- Hozalski R M, Zhang L, Arnold W A, 2001. Reduction of haloacetic acids by Fe: Implications for treatment and fate. *Environ Sci Technol*, 35(11): 2258–2263.
- Kim Y H, Carraway E R, 2000. Dechlorination of pentachlorophenol by zero valent iron and modified zero valent irons. *Environ Sci Technol*, 34(10): 2014–2017.
- Lenga R E, Votoupal K L, 1993. The Sigma-Aldrich Library of Regulatory and Safety Data. Wisconsin, Sigma-Aldrich Corp.
- Lien H L, Zhang W X, 1999. Transformation of chlorinated methanes by nanoscale iron particles. *J Environ Eng-ASCE*, 125(11): 1042–1047.
- Lien H L, Zhang W X, 2001. Nanoscale iron particles for complete reduction of chlorinated ethenes. *Colloids and Surfaces A: Physicochem and Eng Asp*, 191(1-2): 97–105.
- Liu Y H, Quan X, Chen J W, Yang F L, Chen G H, 2000. Quantitative structure-properties relationship for catalytic dechlorination of chlorinated aromatic hydrocarbons. *Jour-*

- nal of Dalian Univ of Technol, 40(3): 301-304.
- Liu Y Q, Lowry G V, 2006. Effect of particle age (Fe⁰ content) and solution pH on NZVI reactivity: H₂ evolution and TCE dechlorination. *Environ Sci Technol*, 40(19): 6085–6090.
- Liu Y Q, Phenrat T, Lowry G V, 2007. Effect of TCE concentration and dissolved groundwater solutes on NZVI-promoted TCE dechlorination and H₂ evolution. *Environ Sci Technol*, 41(22): 7881–7887.
- Lo I M C, Lam C S C, Lai K C K, 2006. Hardness and carbonate effects on the reactivity of zero-valent iron for Cr(VI) removal. *Water Research*, 40(3): 595–605.
- McRae B M, LaPara T M, Hozalski R M, 2004. Biodegradation of haloacetic acids by bacterial enrichment cultures. *Chemosphere*, 55(6): 915–925.
- Muftikian R, Fernando Q, Korte N, 1995. A method for the rapid dechlorination of low molecular weight chlorinated hydrocarbons in water. *Water Research*, 29(10): 2434–2439.
- Quan X, Liu H J, Yang F L, Xue D M, Zhao Y Z, 1998. Dechlorination of three polyhchlorinated hydrocarbons in water using bimetallic systems. *China Environ Sci*, 18(4): 333–336.
- Wei H, Li K B, Tong S P, Liu W P, 2004. Reductive dechlorination atrazine by Ni/Fe bimetallic particles. *Environ Sci*, 25(1): 154–157.
- Wei J R, Wang Z G, 2004. Study advance and control measure on disinfection by-products in drinking water. J Hyg Res, 33(1): 115–118.
- Wu D L, Ma L M, Wang Z, Xu W Y, 2004. A study of factors affecting variation of pH values during dechlorination of chloroorganics. *Industrial Water and Wastewater*, 35(5): 43–46.
- Xu J, Bhattacharyya D, 2007. Fe/Pd nanoparticle immobilization in microfiltration membrane pores: Synthesis, characterization, and application in the dechlorination of polychlorinated biphenyls. *Ind Eng Chem Res*, 46(8): 2348–2359.
- Xu X H, Jin J, Wei J J, Wang D H, 2004. Dechlorination mechanism and kinetic of 2,4-DCP by nanoscale Pd/Fe system. *Acta Scientiae Circumstantiae*, 24(4): 561–567.
- Xu Y, Zhang W X, 2000. Subcolloidal Fe/Ag particles for reductive dehalogenation of chlorinated benzenes. *Ind Eng Chem Res*, 39(7): 2238–2244.
- Yu P, Welander T, 1995. Growth of an aerobic bacterium with trichloroacetic acid as the sole source of energy and carbon. *Appl Microbiol Biotechnol*, 42(5): 769–774.
- Zhang X J, Li S, 2000. Halo acetic acids as an indicator of the total carcinogentic risk of disinfection by-products. *Water and Wastewater Eng*, 26(8): 1–6.
- Zhou H Y, Xu X H, Wang D H, 2003. Catalytic dechlorination of chlorobenzene in water by Pd/Fe. *Journal of Zhejiang University (Engineering Science)*, 37(3): 345–349.

