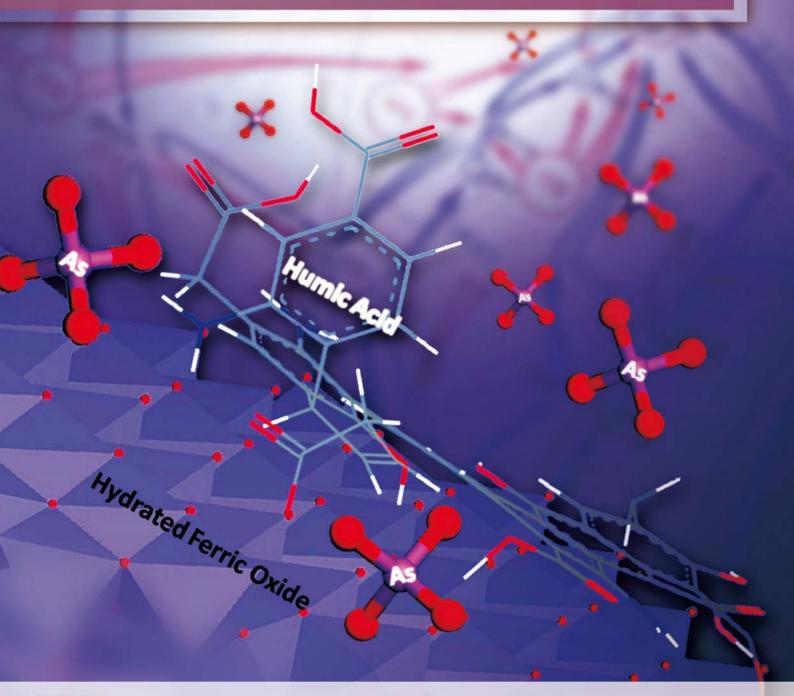
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CONTENTS

Aquatic environment

Removal of total cyanide in coking wastewater during a coagulation process: Significance of organic polymers	
Jian Shen, He Zhao, Hongbin Cao, Yi Zhang, Yongsheng Chen · · · · · · · · · · · · · · · · · · ·	231
Removal of arsenate with hydrous ferric oxide coprecipitation: Effect of humic acid	
Jingjing Du, Chuanyong Jing, Jinming Duan, Yongli Zhang, Shan Hu	240
Arsenic removal from groundwater by acclimated sludge under autohydrogenotrophic conditions	
Siqing Xia, Shuang Shen, Xiaoyin Xu, Jun Liang, Lijie Zhou·····	248
Characteristics of greenhouse gas emission in three full-scale wastewater treatment processes	
Xu Yan, Lin Li, Junxin Liu	256
Effect of temperature on anoxic metabolism of nitrites to nitrous oxide by polyphosphate accumulating organisms	
Zhijia Miao, Wei Zeng, Shuying Wang, Yongzhen Peng, Guihua Cao, Dongchen Weng, Guisong Xue, Qing Yang	264
Efficacy of two chemical coagulants and three different filtration media on removal of Aspergillus flavus from surface water	
Hamid Mohammad Al-Gabr, Tianling Zheng, Xin Yu	274
Beyond hypoxia: Occurrence and characteristics of black blooms due to the decomposition of the submerged plant	
Potamogeton crispus in a shallow lake	
Qiushi Shen, Qilin Zhou, Jingge Shang, Shiguang Shao, Lei Zhang, Chengxin Fan · · · · · · · · · · · · · · · · · · ·	281
Spatial and temporal variations of two cyanobacteria in the mesotrophic Miyun reservoir, China	
Ming Su, Jianwei Yu, Shenling Pan, Wei An, Min Yang·····	289
Quantification of viable bacteria in wastewater treatment plants by using propidium monoazide combined with quantitative PCR (PMA-qPCR)	
Dan Li, Tiezheng Tong, Siyu Zeng, Yiwen Lin, Shuxu Wu, Miao He · · · · · · · · · · · · · · · · · ·	299
Antimony(V) removal from water by hydrated ferric oxides supported by calcite sand and polymeric anion exchanger	
Yangyang Miao, Feichao Han, Bingcai Pan, Yingjie Niu, Guangze Nie, Lu Lv	307
A comparison on the phytoremediation ability of triazophos by different macrophytes	
Zhu Li, Huiping Xiao, Shuiping Cheng, Liping Zhang, Xiaolong Xie, Zhenbin Wu ·····	315
Biostability in distribution systems in one city in southern China: Characteristics, modeling and control strategy	
Pinpin Lu, Xiaojian Zhang, Chiqian Zhang, Zhangbin Niu, Shuguang Xie, Chao Chen · · · · · · · · · · · · · · · · · · ·	323
A4 1	
Atmospheric environment	
Characteristics of ozone and ozone precursors (VOCs and NOx) around a petroleum refinery in Beijing, China	
Wei Wei, Shuiyuan Cheng, Guohao Li, Gang Wang, Haiyang Wang	332
Identification of sources of lead in the atmosphere by chemical speciation using X-ray absorption near-edge structure (XANES) spectroscopy	
Kohei Sakata, Aya Sakaguchi, Masaharu Tanimizu, Yuichi Takaku, Yuka Yokoyama, Yoshio Takahashi	343
Online monitoring of water-soluble ionic composition of PM ₁₀ during early summer over Lanzhou City	
Jin Fan, Xiaoying Yue, Yi Jing, Qiang Chen, Shigong Wang	353
Effect of traffic restriction on atmospheric particle concentrations and their size distributions in urban Lanzhou, Northwestern China	
Suping Zhao, Ye Yu, Na Liu, Jianjun He, Jinbei Chen · · · · · · · · · · · · · · · · · · ·	362
Environmental health and toxicology	
A review on completing arsenic biogeochemical cycle: Microbial volatilization of arsines in environment	
Peipei Wang, Guoxin Sun, Yan Jia, Andrew A Meharg, Yongguan Zhu	371
Alginate modifies the physiological impact of CeO ₂ nanoparticles in corn seedlings cultivated in soil	
Lijuan Zhao, Jose R. Peralta-Videa, Bo Peng, Susmita Bandyopadhyay, Baltazar Corral-Diaz, Pedro Osuna-Avila,	
Milka O. Montes, Arturo A. Keller, Jorge L. Gardea-Torresdey····	382
Humification characterization of biochar and its potential as a composting amendment	
Jining Zhang, Fan Lü, Chenghao Luo, Liming Shao, Pinjing He·····	390
Immigrant Pantoea agglomerans embedded within indigenous microbial aggregates: A novel spatial distribution of epiphytic bacteria	
Qing Yu, Anzhou Ma, Mengmeng Cui, Xuliang Zhuang, Guoqiang Zhuang ······	398
Remediation of nutrient-rich waters using the terrestrial plant, Pandanus amaryllifolius Roxb.	
Han Ping, Prakash Kumar, Bee-Lian Ong	404

Construction of a dual fluorescence whole-cell biosensor to detect N-acyl homoserine lactones
Xuemei Deng, Guoqiang Zhuang, Anzhou Ma, Qing Yu, Xuliang Zhuang·······415
Digestion performance and microbial community in full-scale methane fermentation of stillage from sweet potato-shochu production
Tsutomu Kobayashi, Yueqin Tang, Toyoshi Urakami, Shigeru Morimura, Kenji Kida······423
Health risk assessment of dietary exposure to polycyclic aromatic hydrocarbons in Taiyuan, China
Jing Nie, Jing Shi, Xiaoli Duan, Beibei Wang, Nan Huang, Xiuge Zhao
Acute toxicity formation potential of benzophenone-type UV filters in chlorination disinfection process
Qi Liu, Zhenbin Chen, Dongbin Wei, Yuguo Du · · · · · · · · · · · · · · · · · ·
Exposure measurement, risk assessment and source identification for exposure of traffic assistants to particle-bound PAHs in Tianjin, China
Xiaodan Xue, Yan You, Jianhui Wu, Bin Han, Zhipeng Bai, Naijun Tang, Liwen Zhang······448
Engineers and a stability and materials
Environmental catalysis and materials
Environmental catalysis and materials Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light
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Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light Ashok Kumar Chakraborty, Md Emran Hossain, Md Masudur Rhaman, K M A Sobahan
Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light Ashok Kumar Chakraborty, Md Emran Hossain, Md Masudur Rhaman, K M A Sobahan
Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light Ashok Kumar Chakraborty, Md Emran Hossain, Md Masudur Rhaman, K M A Sobahan
Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light Ashok Kumar Chakraborty, Md Emran Hossain, Md Masudur Rhaman, K M A Sobahan
Fabrication of Bi ₂ O ₃ /TiO ₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light Ashok Kumar Chakraborty, Md Emran Hossain, Md Masudur Rhaman, K M A Sobahan



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Fabrication of Bi₂O₃/TiO₂ nanocomposites and their applications to the degradation of pollutants in air and water under visible-light

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ABSTRACT

A nanoheterojunction composite photocatalyst Bi_2O_3/TiO_2 working under visible-light ($\lambda \geqslant 420$ nm) was prepared by combining two semiconductors Bi_2O_3 and TiO_2 varying the Bi_2O_3/TiO_2 molar ratio. Maleic acid was employed as an organic binder to unite Bi_2O_3 and TiO_2 nanoparticles. The SEM, TEM, XRD and diffuse reflectance spectra were utilized to characterize the prepared Bi_2O_3/TiO_2 nanoheterojunction. The nanocomposite exhibited unusual high photocatalytic activity in decomposing 2-propanol in gas phase and phenol in aqueous phase and, evolution of CO_2 under visible light irradiation while the end members exhibited low photocatalytic activity. The composite was optimized to 5 mol% Bi_2O_3/TiO_2 . The remarkable high photocatalytic efficiency originates from the unique relative energy band position of Bi_2O_3 and TiO_2 as well as the absorption of visible light by Bi_2O_3 .

Introduction

Photocatalytic degradation of organic pollutants on semi-conductor surfaces has attracted much attention as a 'green' technique. TiO_2 , commonly regarded as one of the most active and stable photocatalysts, has long been investigated for environmental applications (Chen and Mao, 2007; Fujishima et al., 2000; Kubacka et al., 2012). However, the band gap of TiO_2 is 3.2 eV which requires the excitation wavelength longer than 387.5 nm. It is a major disadvantage of TiO_2 using as a photocatalyst working under visible-light ($\lambda \geqslant 420$ nm). The high rate of electronhole recombination often results in a low quantum yield and poor efficiency of photocatalytic reactions. Therefore, the development of photocatalysts with a high activity under visible-light irradiation is required. However, an efficient photocatalytic process over a semiconductor de-

mands the high mobility for photoinduced electron-hole separation and for their transportation in crystal lattice, which would lower the probability for electron-hole recombination. Considerable efforts have been made to dope TiO₂ with various metals (In et al., 2007; Wang et al., 2000; Song et al., 2012) as well as non-metals so as to make it photoactive under visible light (Anpo and Takeuchi, 2003; Ho et al., 2012; Sato, 1986). Combination of semiconductors is considered as an effective way to improve the photostimulated electron-hole separation and effectively inhibits their recombination. The major characteristic of this technique is to assemble a heterojunction interface between wide and narrow band gap semiconductors with matching energy band potentials. In this way, electric field assisted transportation of charges from one particle to the other through interfaces is favorable for the electronhole separations in the composite materials, and thus the electron and hole could move to the surface of the semiconductors. The extensive research published on this composite system were mostly focused on TiO2 based pho-



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tocatalysts, such as WO₃/TiO₂, In₂O₃/TiO₂, SiO₂/TiO₂, MgO/TiO₂, Fe₂O₃/TiO₂, Bi₂O₃/TiO₂, FeTiO₃/TiO₂ and so on (Bian et al., 2008; Chakraborty and Kebede, 2012; Ding et al., 2000; Gondal et al., 2004; Gao et al., 2008; Kim et al., 2009; Naeem and Ouyang, 2013; Song et al., 2001).

Bi₂O₃, with a band gap of 2.8 eV is known as p-type semiconductor (Hameed et al., 2008; Lin et al., 2007; Xu and Schoonen, 2000), and it has widely been used in gas sensors, optical coatings and solid oxide fuel cells. Moreover, Bi₂O₃ has proved to be a good photocatalyst for water splitting and pollutant decomposing under visible-light irradiation (Adamian et al., 1996; Fan et al., 2005; Fruth et al., 2005; Leontie et al., 2002; Yang et al. 2004; Zhou et al., 2009). Bi₂O₃ semiconductor possesses good electron (e⁻) conduction ability. On the other hand, TiO2 has a loosely packed structure as well as higher degree of openness (Lin et al., 2006), which favors the hole (h⁺) transport in the crystal lattice by the available displacement of the oxygen atoms through the strong vibration model (associated with O⁻). Thus, in such a heterojunction system, one component, Bi₂O₃ of the composite with fair electron (e⁻) conductivity cooperates with a semiconducting material, TiO₂ with an open structure and the fair mobility for the hole (h⁺) conduction. Hence, when Bi₂O₃/TiO₂ heterojunction is formed, the inner electric field will be established in the interface (Lin et al., 2007). The photogenerated electron-hole pairs will be separated effectively by the inner electric field, and the photocatalytic activity is enhanced. Such a phenomenon gives us an implication for a design scheme for the efficient heterojunction photocatalysts between Bi₂O₃ and TiO₂.

In the present study, we have developed a heterojunction Bi_2O_3/TiO_2 system utilizing maleic acid as an organic binder. The role of maleic acid is to bind Bi_2O_3 and TiO_2 using two end carboxylic functional groups. The prepared Bi_2O_3/TiO_2 nanoheterojunction was applied to the photocatalytic degradation of 2-propanol in gas phase and phenol in aqueous phase under visible-light irradiation. The composite Bi_2O_3/TiO_2 demonstrated enhanced photocatalytic efficiency compare to that of Bi_2O_3 and Degussa P25 under-visible light irradiation.

1 Materials and methods

1.1 Preparation of Bi₂O₃/TiO₂ composite

Bi₂O₃/TiO₂ composite photocatalyst was prepared by utilizing maleic acid. Degussa P25 was chosen as the standard TiO₂ nanoparticle. In a typical preparation for 5/95 Bi₂O₃/TiO₂ (the composite consisting 5 mol% Bi₂O₃ and 95 mol% TiO₂) heterojunction photocatalyst, 0.3070g of Bi₂O₃ (99.9% Aldrich) was suspended in 40 mL of absolute ethanol. A 0.1987 g of maleic acid (99.99% Aldrich) dissolved in absolute ethanol was added to this

suspension. Afterward, 1 g of TiO_2 nanoparticle was added to the above suspension and stirred vigorously for 6 hr at ambient condition. Subsequently, the suspension was centrifuged and Bi_2O_3/TiO_2 composite was washed several times with ethanol to remove unreacted maleic acid. The composite was dried at $60^{\circ}C$ in an oven overnight. After that, the Bi_2O_3/TiO_2 composite was annealed at $300^{\circ}C$ for 3 hr and used as catalyst for photocatalytic degradation reactions without further treatment.

1.2 Characterization

The crystallographic phase of prepared samples was characterized by a Rigaku Multiflex diffractometer with monochromated high-intensity Cu Kα radiation. XRD scanning was performed under ambient conditions over the 2θ region of 20° – 60° at a rate of 2° /min (40 kV, 20 mA). The morphologies of the prepared powders were examined by a scanning electron microscope (Hitachi S-4500). The surface morphology of the Bi₂O₃/TiO₂ composite was analyzed by a transmission electron microscope (TEM, Philips CM30) operating at 250 kV. For this measurement, 1 mg of Bi₂O₃/TiO₂ composite was dispersed in 50 mL of ethanol, and a drop of the suspension was then spread on a holey amorphous carbon film deposited on a Ni grid (JEOL Ltd.). Diffuse reflectance spectra were also obtained for the dry-pressed disk sample using a UV-Visible spectrometer (Perkin-Elmer Lambda 40).

1.3 Measurement of photocatalytic efficiency

The composite Bi₂O₃/TiO₂ films were tested as photocatalysts for the decomposition of 2-propanol in gas phase. An aqueous suspension containing 4×10⁻⁴ mol of Bi₂O₃/TiO₂ composite were spread on a 2.5×2.5 Pyrex glass in the film form, which was subsequently dried at 50°C for 2 hr. The gas reactor system used for this photocatalytic reaction is described elsewhere (Kwon et al., 2000). The reaction cell was placed closed to GC and it was connected to GC of which one window is open and the cutoff filter was set in front of this window face of the reaction cell to ensure the desired irradiation condition. The volume of the reaction cell was 200 mL. The film was located at the center of this gastight reactor. After evacuation of the reactor, 0.08 µL of 2-propanol mixed in 1.6 µL of water was injected into the reactor. The total pressure of the reactor was then controlled to 750 Torr by addition of oxygen gas. Under these conditions, 2-propanol and H₂O remained in the vapor phase. The reactor was irradiated from a 300 W Xenon lamp equipped with a water filter and UV filter (< 420 nm, Oriel) to cut off infrared and UV component, respectively. On the surface of the photocatalytic film, the overall light intensity was 0.1 W/cm². After irradiation for a certain period of time, 0.5 mL of the gas in the reactor was automatically picked up and sent to a gas chromatograph (Agilent Technologies, Model 6890N). The whole photocatalytic reaction was batch process. The

evolved CO_2 and remnant 2-propanol were detected by a methanizer installed between the GC column outlet and the FID detector.

The initial concentration of gaseous 2-propanol in the reactor was kept to 117 ppm in volume (ppmV). Thus the ultimate concentration of evolved CO₂ will be 351 ppmV when the whole 2-propanol is completely decomposed, as shown in the following Reaction:

$$2(CH_3)_2CHOH(g) + 9O_2(g) \longrightarrow 6CO_2(g) + 8H_2O \qquad (1)$$

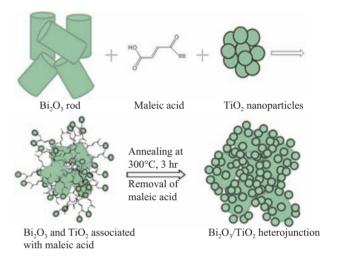
The photocatalytic activity of the prepared nanocomposite sample was evaluated by the decomposition of phenol in aqueous solution. Each photocatalyst was dispersed in water and then mixed with phenol aqueous solution in darkness. Total volume of the aqueous solution for the photocatalytic reaction was 50 mL, and the concentration of phenol was $1.00\times10^{-4} \text{ mol/L}$, and $1.00\times10^{-4} \text{ mol}$ of photocatalytic sample was suspended in this solution. The prepared solution was then irradiated by a 300-W Xe lamp with an UV cutoff filter (< 420 nm, Oriel) and a water filter to remove IR. The concentration of residual phenol after irradiation was evaluated from the intensity of the characteristic absorption peak.

Prior to photocatalytic reactions, the reaction medium were magnetically stirred for 30 min in the dark, and the concentration of the pollutants were monitored. The concentration of pollutants did not change after stirring for 30 min which indicates that 30 min is enough to reach the adsorption equilibrium of organics. Blank experiments were also performed at the same experimental condition except photocatalyst.

2 Results and discussion

In preparing Bi_2O_3/TiO_2 heterojunction, maleic acid has been considered. The function of maleic acid is to chemically combine Bi_2O_3 and TiO_2 nanoparticles utilizing the two carboxylic functional groups as shown in **Scheme 1**.

Thus, the Bi_2O_3 and TiO_2 nanoparticles move towards each other and, Bi_2O_3 and TiO_2 could make an intimate contact which is vital for interfacial charge transfer between the two semiconductors. However, the composite was annealed for 300° C to eliminate maleic acid. Hence, the maleic could not influence the photocatalytic activity. The SEM images of pure Bi_2O_3 rod and 5/95 Bi_2O_3/TiO_2 composite structure are shown in **Fig. 1**. Images in **Fig. 1a**, **b** reveal the Bi_2O_3 exhibits a rod like morphology, where



 $\label{eq:Scheme 1} \begin{tabular}{ll} Scheme 1 & Schematic illustration of the formation mechanism of Bi$_2O_3/$\\ TiO_2 & heterojunction. \end{tabular}$

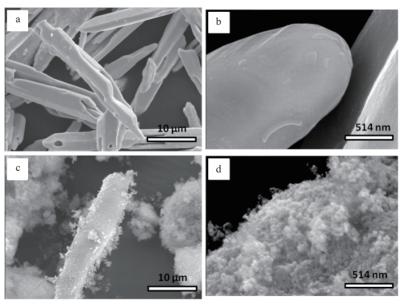


Fig. 1 SEM images of Bi₂O₃ (a, b) and Bi₂O₃/TiO₂ (c, d) nanocomposite calcined at 300°C for 3 hr.

the length and diameter of the rods varied over the ranges of 4–33 and 0.4–3.0 µm, respectively. Maleic acid plays a crucial role in combining Bi_2O_3 and TiO_2 nanoparticles. The surface of Bi_2O_3 is fine and smooth as shown in **Fig. 1a, b**. The surface morphology of Bi_2O_3/TiO_2 composite is shown in **Fig. 1c, d**. The surface of Bi_2O_3/TiO_2 composite is completely different from pure Bi_2O_3 and uneven which is shown in **Fig. 1c**. To reveal the detail structure of the composite, **Fig. 1c** was further magnified and shown in **Fig. 1d**. **Figure 1d** clearly demonstrates the surface structure of Bi_2O_3/TiO_2 composite that the surface of heterojunction is rough and uniformly covered with TiO_2 nanoparticles.

In order to investigate the surface structure and the interface of the composite very closely, the sample of 5/95 Bi₂O₃/TiO₂ composite calcined at 300°C was chosen to take TEM and HRTEM characterization. **Figure 2** gives an overview of the typical TEM image of the 5/95 Bi₂O₃/TiO₂ composite. Nanoparticles of Degussa P25 with mean size of about 20–30 nm can be observed over the entire surface of Bi₂O₃. The high resolution TEM image of the magnified view of the composite is shown in **Fig. 2c, d.** A clear and sharp interface can be found in these

images. It is indicated that the two components of the composite are contacted well after the preparation and subsequent heat treatment. The high crystalline quality and the sharp interface between Bi₂O₃ and TiO₂ nanoparticles are advantageous for the separation of the photogenerated charge carriers.

Figure 3 shows the XRD patterns of TiO₂ nanoparticles, Bi₂O₃ and several empositions of Bi₂O₃/TiO₂ composite samples. The diffraction peaks at 21.82°, 27.48°, 28.12°, 32.61°, 46.49°, 52.56° and 54.99° were indexed to the (020), (120), (012), (211), (223), (321) and (241) planes, respectively to the pure Bi₂O₃ (JCPDS #72-0398). As shown in **Fig. 3** that the diffraction planes of (101), (004), (200) and (211) clearly indicate the presence of anatase and that of (110) and (101) planes demonstrate the existence of rutile phase of TiO2 nanopartiles. Hence, Degussa P25 contains anatase and rutile phase. It is worth mentioning that the XRD patterns of the composites Bi₂O₃/TiO₂ showed the diffraction peaks of phases of Bi₂O₃ and TiO₂ without any other impurity phases suggesting that there was no sufficient chemical reaction took place among the component Bi₂O₃ and TiO₂ of the composite during the preparation of heterojunction and subsequent heat-

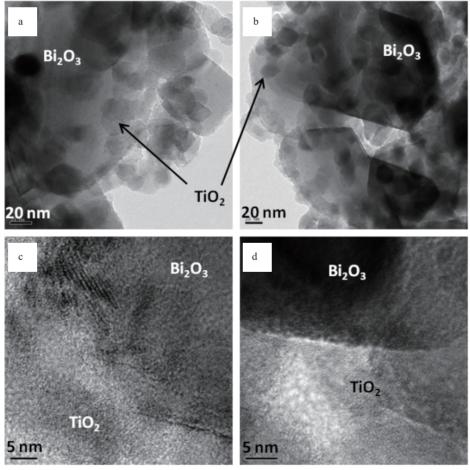


Fig. 2 TEM (a, b) and HRTEM (b, c) images of 5/95 Bi₂O₃/TiO₂ heterojunction calcined at 300°C for 3 hr.

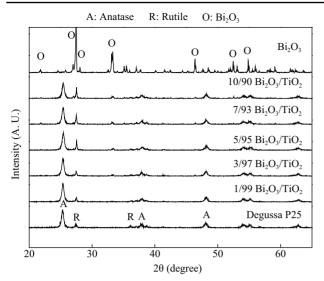


Fig. 3 X-ray diffraction pattern of TiO_2 , Bi_2O_3 and Bi_2O_3/TiO_2 composite synthesized at $300^{\circ}C$ for 3 hr with several composition of Bi_2O_3 and TiO_2 .

treatment at 300° C. As shown in **Fig. 3** that the intensity of the major peak of Bi_2O_3 of (120) plane at 27.48° gradually increases with increasing amount of Bi_2O_3 from 0% to 100% in the composite Bi_2O_3/TiO_2 . However, the annealing temperature 300° C is sufficiently enough to make a tight contact between Bi_2O_3 and TiO_2 nanoparticles as shown in **Fig. 2c** and **d**. The well contact will facilitate interparticle electron transfer between Bi_2O_3 and TiO_2 .

The UV-Vis diffuse reflectance spectra of the asprepared composite powders are shown in **Fig. 4.** The band gaps of Bi₂O₃ and TiO₂ were reported to be 2.8 and 3.2 eV, respectively (Hameed et al., 2008; Xu and Schoonen, 2000). The optical absorptions of Bi₂O₃/TiO₂ composite powders start at about 450 nm, corresponding to the absorption edge of Bi₂O₃. The second absorption edge at about 387 nm can be attributed to the TiO₂ nanoparticles. It is mentionable from **Fig. 4** that with increasing the

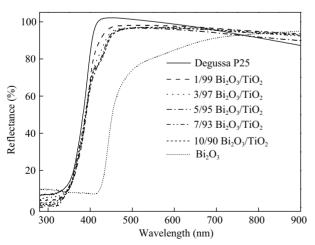


Fig. 4 UV-Visible diffuse reflectance spectra of Bi_2O_3/TiO_2 composite, Bi_2O_3 and TiO_2 nanoparticle.

component of the Bi_2O_3 in the composite Bi_2O_3/TiO_2 system, the absorption in visible-light of the composite powders also increased. These results revealed that the composite powders could absorb the photon in the visible region of the solar spectrum.

The dependence of photocatalytic activity on the composition of the composite Bi_2O_3/TiO_2 system under visible light ($\lambda \geqslant 420$ nm) irradiation is represented in **Fig. 5**. The catalyst samples were evaluated for the degradation of 2-propanol and the production of CO_2 in gas phase and phenol in aqueous phase under visible light irradiation at room temperature. It is generally accepted that IP is first decomposed to acetone and finally to CO_2 . Thus, we measured the photocatalytic decomposition behavior of IP first. **Figure 5a** shows the visible light ($\lambda \geqslant 420$ nm) activity of the

 Bi_2O_3 , TiO_2 and Bi_2O_3/TiO_2 composite photocatalyst for the decomposition of 2-propanol. Bi_2O_3 , TiO_2 exhibited negligible activity, while Bi_2O_3/TiO_2 in several compositions showed higher visible light induced photocatalytic activity. When photocatalytic reaction proceeded for 2 hr, the decomposition degrees of gaseous 2-propanol was 75% with 5/95 Bi_2O_3/TiO_2 heterojunction as shown in **Fig. 5a**.

The photocatalytic efficiency of the composite sample was investigated for the evolution of CO_2 . **Figure 5b** displays the photocatalytic behavior of Bi_2O_3/TiO_2 composite samples including the component of the composite Bi_2O_3 and Degussa P25 to compare the efficiency. The composite Bi_2O_3/TiO_2 samples with a wide range of composition demonstrated significant production of CO_2 . As that of the decomposition of gaseous 2-propanol, 5/95 Bi_2O_3/TiO_2 demonstrated the highest activity among the composition for the production CO_2 . After 2 hr of visible light ($\lambda \ge 420$ nm) irradiation, the detected CO_2 yield was reached to a stable value of 21 ppmV, whereas both the bare Bi_2O_3 and Degussa P25 showed only 2.1 and 2.9 ppmV, respectively.

We evaluated the photocatalytic activities of these catalysts in aqueous phase for the decomposition of phenol under visible light ($\lambda \ge 420$ nm). The remnant phenol after the irradiation of visible light was analyzed from its characteristic absorption peak detected by UV-vis spectroscopy. Figure 5c shows decomposition rates of phenol with several Bi₂O₃/TiO₂ composition, Bi₂O₃, and P25 samples as a function of irradiation time. It was found that the photocatalytic activity of the composite was increased with increasing Bi₂O₃. Similar results were obtained as that of decomposition of gaseous 2-propanol, 5/95 Bi₂O₃/TiO₂ showed the highest photocatalytic efficiency in decomposing phenol in aqueous phase. The decomposition of phenol was 69% with this composition in 2 hr of visible light irradiation. Blank experiments as shown in Fig. 5 that the photodecomposition of the organics in gas and aqueous phases was not took place.

Coupling of TiO₂ with other narrow band gap semi-

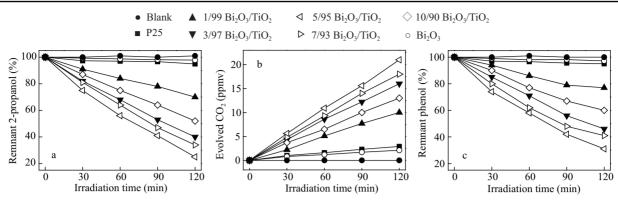
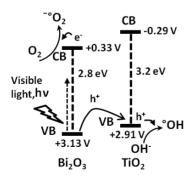


Fig. 5 Photocatalytic activity of TiO_2 , Bi_2O_3 and Bi_2O_3/TiO_2 composite for (a) degradation of 2-propanol, (b) evolution of CO_2 and (c) decomposition of phenol under visible light ($\lambda \ge 420$ nm) irradiation.

conductors has been extensively studied to promote the separation of photogenerated charge-carriers and/or to extend the absorption of photon up to visible-region as well as to enhance the photocatalytic behavior in decomposing organic pollutants. The present investigated heterojunction Bi₂O₃/TiO₂ exhibited extraordinary high photocatalytic efficiency but both the components Bi₂O₃ and TiO₂ demonstrated very low photocatalytic activity under visible-light irradiation. Presumably, the unusual high photocatalytic activity of Bi₂O₃/TiO₂ originates from the unique relative energy band positions of these two semiconductors (Chai et al., 2009; Chakraborty and Kebede, 2012; Chakraborty et al., 2012; Gao et al., 2008, 2010; Kim et al., 2009) and the absorption of visible light by Bi₂O₃. As shown in Fig. 6, the VB of Bi₂O₃ is located below than that of TiO₂ by 0.22V, while CB of Bi₂O₃ is positioned below than that of the TiO₂ by 0.62V (Hameed et al., 2008; Xu and Schoonen, 2000). During visible-light irradiation, Bi₂O₃ absorbs visible-light as the band gap of Bi₂O₃ is 2.8 eV and the electrons (e⁻) in the VB of Bi₂O₃ would be excited to its CB and the VB of Bi₂O₃ become partially vacant, and the generated holes (h⁺) in the VB of Bi₂O₃ can be transferred to VB of TiO₂, since the VB of TiO₂ is located higher than the VB of Bi₂O₃. Thus, the holes (h⁺) in the VB of TiO₂ can be utilized for various oxidation reactions.

The CB potential of Bi₂O₃ is +0.33V. Hence, the reduc-



 $\label{eq:Fig.6} \textbf{Fig. 6} \qquad \text{Proposed mechanism for the photocatalytic degradation of organic pollutants over Bi_2O_3/TiO_2 heterojunction photocatalyst.}$

tion of oxygen utilizing single electron is quite difficult (Reactions (2) and (3)). Thus, the CB electrons of Bi_2O_3 would be consumed through multi-electron process (Abe et al., 2008; Liu et al., 2009; Rawal et al., 2012) according to following reactions (Reactions (4) and (5)).

$$O_2 + e^- = O_2^-;$$
 $E^0 = -0.284 \text{ V (vs. NHE)}$ (2)

$$O_2 + H^+ + e^- = HO_2;$$
 $E^0 = -0.046 \text{ V (vs. NHE)}$ (3)

$$O_2 + 2H^+ + 2e^- = H_2O_2$$
; $E^0 = +0.682 \text{ V}$ (vs. NHE) (4)

$$O_2 + 4H^+ + 4e^- = 2H_2O$$
; $E^0 = +1.23 \text{ V (vs. NHE)}$ (5)

By the inter-semiconductor hole-transfer mechanism, the photogenerated charge-carriers can be effectively separated so that this heterojunction can utilize the visible-light in mineralizing organic pollutants. In this photocatalytic system, TiO₂ works as principal photocatalyst, while the role of Bi₂O₃ is a sensitizer absorbing visible-light to sensitize TiO₂. We expect that this synthesis process can also be extended to the development of other heterojunction type composite photocatalysts working under sunlight to remove organic pollutants.

3 Conclusions

A series of Bi_2O_3/TiO_2 heterocomposite photocatalyst have been prepared by utilizing maleic acid. It was observed that the composite could absorb energy up to visible-region. The new composite demonstrated notably higher photocatalytic activity for the degradation of 2-propanol in gas phase and phenol in aqueous phase and, evolution of CO_2 than Bi_2O_3 and TiO_2 nanoparticles under visible light ($\lambda \geq 420$ nm) irradiation. After 2 hr of visible light ($\lambda \geq 420$ nm) irradiation, the decomposition of 2-propanol and phenol is 75% and 69%, respectively and, the evolution of CO_2 is 21 ppmV. The evolved CO_2 is about 7.5 times that of Degussa P25. TiO_2 works as a principal photocatalyst while Bi_2O_3 acts as the photosensitizer absorbing visible light in the Bi_2O_3/TiO_2 composite.

The extraordinary high photocatalytic efficiency of the Bi_2O_3/TiO_2 composite originates from the unique relative energy band position of the component semiconductors. The VB level of Bi_2O_3 (+3.13V vs. NHE) is located lower than that of TiO_2 (+2.91V vs. NHE). The visible light generated hole (h⁺) in the VB of Bi_2O_3 can be transferred to the VB of TiO_2 . Therefore, with the irradiation of visible light the holes (h⁺) in the VB of TiO_2 can induce the complete decomposition of organic pollutants.

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